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1 **Phthalates and Bisphenols Migration in Mexican Food Cans and Plastic Food**
2 **Containers**

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21 Running Head: Phthalates and bisphenols in food containers

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33 **Abstract:** The presence of endocrine disruptors bisphenol-A, bisphenol-A-dimethacrylate,
34 bisphenol-A-diglycidyl-ether, phthalic-acid, dibutyl-phthalate, diethyl-phthalate and
35 dioctyl-phthalate was determined in vegetable cans, baby bottles and microwaveable
36 containers from the Mexican market. Gas-Chromatography-Mass-Spectrometry was used
37 for the identification and High-Performance-Liquid-Chromatography with UV/Visible
38 light and fluorescence detectors was used for the quantification. Endocrine disruptors were
39 found in all samples. PA and DOP were the substances most commonly found, and
40 maximum concentrations were 9.549 $\mu\text{g}/\text{kg}$ and 0.664 $\mu\text{g}/\text{kg}$, respectively from a jalapeno
41 peppers can. Bisphenol A, phthalic-acid, bisphenol-A-dimethacrylate and dibutyl-phthalate
42 were found in baby bottles and microwaveable containers.

43
44 **Keywords:** Baby bottles, cans, endocrine disruptor, plastic.

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46
47 Synthetic polymers known as epoxy resins contain among their components, bisphenol A
48 (BPA) or its derivatives, which are used to improve the durability of inner varnishes of
49 food cans and in the manufacturing of different kinds of food containers, such as baby
50 bottles, microwaveable containers and plastic dairy containers. Other molecules such as
51 phthalates are used to increase the flexibility of plastics and in the manufacture of teething
52 rings and pacifiers (European Chemicals Bureau 2003). Endocrine disruptors as bisphenols
53 and phthalates have harsh effects on the human reproductive system (Roy et al. 2009). The
54 bisphenols also are associated to endometrial hiperplasia and endometrial cancer in human
55 beings (Hiroi et al. 2004). Several reports have shown the migration of BPA from cans and
56 baby bottles to foods (Cao and Corriveau 2008; Yonekubo et al. 2008). Infants, whose
57 growth and development are highly dependent upon the endocrine system, are particularly
58 vulnerable to endocrine disruptor exposure (Ozaki et al. 2002). However, there are few
59 reports showing the migration of phthalic acid and their esters in food cans or plastic food
60 containers, despite that they also exhibit endocrine disruption activity and may produce
61 hepatocarcinogenesis (Huber 1996; Ozaki et al. 2002).

62 The aim of this work was to determine the migration of phthalates and bisphenols from
63 Mexican vegetable cans, microwaveable containers, baby bottles and plastic food
64 containers, subjected to high temperatures.

65 66 67 **Materials and methods**

68
69 All chemicals were purchased from Sigma-Aldrich with purity of 99.9%. The analyzed
70 samples included 24 cans of vegetables (sweet corn kernels, jalapeno peppers, mushrooms,
71 peas, and peas and carrots), 4 plastic microwave oven containers, 12 plastic (yoghurt or
72 cream) food containers and 6 baby bottles. All of the sample items were bought in
73 supermarkets of San Luis Potosi, Mexico.

74 The extraction of bisphenols and phthalates from vegetable cans was done using water as a
75 food simulant liquid according to the following procedure: 300 mL of water (Milli-Q)
76 were added to each sample item and sterilized at 121°C for 15 min. An aliquot of 75 mL
77 was taken and 20 mL of methanol were added. The aliquot was vigorously agitated for 10
78 min in a decantation funnel, and then 40 mL of organic solvent (see table 1) were added.
79 The aliquot was agitated once again for 10 min and centrifuged at 1200 g for 15 min. The

80 organic phase was taken and concentrated in a rotary evaporator at 40°C and reduced
81 pressure to a final volume of 10 mL. The solvent was further eliminated using a nitrogen
82 stream to a final volume of 1 mL. Then 0.5 mL of sulphuric acid was added and
83 centrifuged at 1200 g for 10 min. The organic phase was aspirated and dried through a
84 stream of gaseous nitrogen. The dry remains were solubilised in 1 mL of methanol and
85 analyzed by HPLC and GC-MS as described below. The determination of disruptors
86 migration from baby bottles and microwave containers was based on the methodology
87 reported by Brede et al. (2003) using a microwave oven (Samsung Co., Malaysia).

88 Compound identification was done with a Gas Chromatograph coupled to a Mass
89 Spectrometry detector Starum 2000 (Varian Instrument, Walnut Creek, CA) with an
90 automatic injector 8200, SPI/1078 (Varian Instrument, Walnut Creek, CA), and using a
91 capillary column DB5-MS (Varian Instrument, Walnut Creek, CA) 30-meters long and
92 0.25 mm diameter. The injector temperature was 250°C. The GC was operated in a split-
93 less mode using helium as the carrier gas at 1 mL/min. The temperature program started at
94 50°C for 3 min, and it was increased by 30°C/min rate until 200°C was reached. The
95 temperature was held at 200°C for 5 min. Then the temperature was increased at 30°C/min
96 to 270°C and it held at 270°C for 8 min. Followed by another temperature increase of
97 30°C/min to 310°C. The temperature was held at 310°C for 2 min. The ionization potential
98 was 70 eV and a scan function with 100-300 *m/z*. Compounds were identified by
99 comparing the mass spectrum with the spectra from the NIST library of the Masslab
100 software (Thermo Quest, Manchester, UK).

101 Analysis by High Performance Liquid Chromatography was performed using a C-18
102 column 4.6 x 150 mm with a particle size of 5 µm of octadecylsilicate (Merck, Darmstadt,
103 Germany) in a HPLC System 600 (Waters) connected to an UV/Vis 2487 detector
104 (Waters) and a fluorescence 2475 detector (Waters). The UV/Vis detector was set at 280
105 nm, and the fluorescence detector had an excitation wavelength of 275 nm and an emission
106 wavelength of 300 nm. The solvents were acetonitrile 50% (phase A) and acetonitrile
107 100% (phase B), at 1 mL/min of flow rate with a gradient 0-100% of B during 20 min, and
108 a temperature of 25°C. The volume of the injected sample was 20 µL. To verify the
109 retention times of analytes, ten samples, dissolved in methanol were injected, and
110 arithmetic means were calculated.

111 Detection and quantification limits were measured by the signal/noise method and using
112 the standard deviation of the response and the slope (Long and Winfordner 1983). Linear
113 response of detectors was determined by straight calibration curves from the analysis of
114 five standards with increasing analyte concentrations (Thompson et al. 2002).

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116

117 **Results and Discussion**

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119 The percentage of endocrine disruptors recovered using hexane, hexane:ethylic ether or
120 chloroform during the extraction protocol is shown in table 1. Chloroform showed the
121 highest recovery percentage with values ranging from 75.0±8.8% to 95.0±5.4% for PA and
122 DBP respectively, whereas hexane had the lowest values. Based on these results
123 chloroform was used in subsequent experiments.

124 The endocrine disruptors were identified by CG-MS as described in materials and
125 methods. Meanwhile the quantitative analysis was carried out by HPLC. The retention

126 times, detection and quantification limits are shown in the table 2. The detection limit for
 127 BPA in this study was less than that reported by others authors (Lim et al. 2009).
 128

129 **Table 1** Percentage of endocrine disruptors recovered after the extraction using different
 130 organic solvents.

Compounds*	Quantity (ng)	Hexane (%)	Hexane: Ethylic ether 1:1 (%)	Chloroform (%)
PA	200.0	12.0±3.4	45.0±5.0	85.0±7.9
	400.0	20.0±2.9	65.0±6.4	88.0±8.7
	600.0	24.0±2.9	66.0±7.9	75.0±8.8
BPA	30.0	-	63.1±4.1	88.4±6.4
	60.0	-	60.2±5.6	88.1±4.1
	100.0	-	62.8±2.1	89.2±4.5
BADGE	3.0	-	54.0±6.6	90.0±2.3
	5.0	-	69.0±7.3	90.0±5.3
	7.0	-	64.0±2.9	92.5±3.2
BisDMA	2.0	-	67.3±4.6	88.0±5.3
	4.0	-	62.2±5.2	80.0±5.5
	6.0	-	60.7±4.5	82.0±3.8
DOP	4.0	22.3±1.8	31.4±2.2	87.5±7.6
	6.0	24.5±2.2	33.2±4.1	83.9±6.5
	8.0	21.7±3.1	30.2±2.9	84.7±5.3
DBP	0.40	45.0±4.9	58.2±4.1	90.0±7.3
	0.60	50.0±4.9	55.2±3.3	95.0±5.4
	0.80	40.0±3.9	52.2±4.2	88.0±5.7
DEP	0.40	33.2±3.1	43.2±2.5	92.0±5.4
	0.60	31.4±3.3	40.8±3.8	92.0±5.3
	0.80	30.4±2.8	40.5±5.2	90.0±5.6

131 *PA: phthalic-acid, BPA: bisphenol-A, BADGE: bisphenol-A-diglycidyl-ether,
 132 BisDMA: bisphenol-A-dimethacrylate, DOP: dioctyl- phthalate, DBP: dibutyl-
 133 phthalate and DEP: diethyl-phthalate.
 134

135 **Table 2** Retention time, detection and quantification limits of the endocrine disruptors
 136 analyzed by HPLC and using the signal/noise method.

Compounds*	Retention time (min.)**	Detection limit ($\times 10^{-4}$ $\mu\text{g}/\text{kg}$)	Quantification limit ($\times 10^{-4}$ $\mu\text{g}/\text{kg}$)
PA	1.60 ± 0.20	1.0	3.0
BPA	3.20 ± 0.01	0.01	0.02
DEP	5.61 ± 0.05	10.0	60.0
BADGE	10.70 ± 0.04	30.0	80.0
BisDMA	11.70 ± 0.05	30.0	90.0
DBP	12.70 ± 0.01	50.0	20.0
DOP	16.50 ± 0.60	1.5	4.0

137 *PA: phthalic-acid, BPA: bisphenol-A, BADGE: bisphenol-A-diglycidyl-ether,
 138 BisDMA: bisphenol-A-dimethacrylate, DOP: dioctyl- phthalate, DBP: dibutyl-

139 phthalate and DEP: diethyl-phthalate. BPA was analyzed using a fluorescence
 140 detector and for the other compounds were analyzed by UV/Vis detector.
 141 **Means and standard deviations were calculated from n=10 independent
 142 experiments.
 143

144 The results for the analysis of the endocrine disruptors in the samples of vegetable cans are
 145 shown in table 3. PA, BPA, DEP, DBP and DOP were found, while BADGE and BisDMA
 146 were not detected. PA and DOP were the compounds most frequently found, and the
 147 maximum concentrations were 9.549 µg/kg and 0.664 µg/kg, respectively in a can of
 148 jalapeno peppers. However, Thomson et al. (2005) found levels of 10 to 29 µg/kg of BPA
 149 in 80 different samples of food cans of New Zealand. BPA concentrations between 29.9
 150 and 80 µg/kg were found in 20 samples of vegetable cans from Brazil, France, Spain,
 151 Turkey and USA (Brotons et al. 1995). Whereas, Kang et al. (2003) found BPA levels of 5
 152 µg/kg in food cans. The migration of bisphenol derivatives obtained in our work, was less
 153 than those reported by Yonekubo et al. (2008) in 38 canned foods bought in Japan.
 154

155 **Table 3** Endocrine disruptors concentration (µg/kg) found in Mexican vegetable cans.

Sample	PA	BPA (x10 ⁻³)	DEP	DBP	DOP
Sweet corn kernel n=6	0.001	3.400	0.023	0.006	0.082
	0.398	0.010	-	-	0.262
	0.339	-	0.030	0.001	-
	0.034	-	0.042	-	0.038
	2.187	-	-	0.008	0.060
Mushrooms n=2	0.099	-	0.035	0.006	0.080
	0.068	2.300	-	-	0.031
Jalapeno peppers n=6	1.697	1.200	-	-	0.060
	9.549	15.900	-	-	0.664
	0.082	-	-	0.023	-
	0.484	-	-	0.018	-
	0.016	0.400	-	-	0.030
Peas n=5	1.558	21.300	-	-	0.060
	0.034	-	0.032	-	0.070
	0.056	-	0.022	0.008	-
	0.018	-	-	0.006	0.090
	0.014	2.500	-	0.007	0.100
Peas and carrots n=5	-	4.700	0.033	-	0.086
	0.037	2.100	0.018	-	0.076
	0.030	-	-	-	0.024
	5.876	1.700	-	-	0.034
	5.172	-	-	-	0.510
	0.013	0.400	-	-	0.071
	0.088	-	0.013	-	0.060

156
 157 The analysis of monomers in the plastic food containers, microwaveable containers and
 158 baby bottles is shown in table 4. PA, BPA, DEP, DBP, DOP, BADGE and BisDMA were

159 detected. PA and DOP were the compounds most frequently found, and the maximum
 160 concentrations were 3.866 and 0.228 $\mu\text{g}/\text{kg}$, respectively. The maximum PA
 161 concentrations were found in a cream container and a yoghurt container, respectively.
 162 BADGE, BisDMA and DBP were not detected as frequently. They were found in baby
 163 bottles, yoghurt and microwaveable containers. The importance of these findings is that
 164 plastic food containers such as yoghurt and the cream are re-used in Mexican households
 165 to keep and to reheat food in microwave ovens, exposing the food to high temperatures
 166 and increasing the risk of migration. Kang and Kondo (2003) found BPA in 40 plastic food
 167 containers (yoghurt, butter, cream, puddings and condensed milk) in a range of 21-43
 168 $\mu\text{g}/\text{kg}$. Brede et al. (2003) analyzed twelve different polycarbonate baby bottles both new
 169 and second hand from Norway using water as food simulant by solid phase extraction,
 170 followed by CG-MS analysis. They found bisphenols migration of 0.23 $\mu\text{g}/\text{L}$ in the new
 171 baby bottles, 8.4 $\mu\text{g}/\text{L}$ in bottles dishwashed 51-times, and 6.7 $\mu\text{g}/\text{L}$ in bottles dishwashed
 172 169-times. Cao and Corriveau (2008) found bisphenols in baby bottles purchased in
 173 Canada, ranging from 228 to 521 $\mu\text{g}/\text{kg}$ after warming them up to 70°C for 6 days.
 174 Maragou et al. (2008) found levels of bisphenols migration from 2.4 to 14.3 $\mu\text{g}/\text{kg}$ in 31
 175 polycarbonate (PC) baby bottles purchased in Greece.

176
 177 **Table 4** Endocrine disruptors concentration ($\mu\text{g}/\text{kg}$) found in food plastic containers
 178 purchased in Mexico.

Sample	PA	BPA	DEP	BADGE	BisDMA	DBP	DOP
Microwaveable n=4	-	0.001	-	-	0.024	0.001	-
	0.023	0.002	-	-	-	-	0.076
	0.577	-	-	-	-	0.003	-
	3.251	0.002	-	-	-	-	0.057
Yoghurt n=7	0.011	0.001	-	0.002	0.147	-	-
	-	-	-	-	0.001	-	0.078
	0.432	-	0.012	-	-	-	0.063
	0.251	-	-	-	-	-	0.215
	2.890	0.002	-	-	-	-	0.107
	0.491	-	0.011	-	-	-	0.041
	0.090	0.001	-	-	-	-	0.228
Cream n=5	1.036	-	0.011	-	-	-	0.200
	0.963	-	-	-	-	-	0.115
	0.019	-	0.015	-	-	-	0.077
	1.330	-	0.012	-	-	-	0.004
	3.866	-	0.096	-	-	-	-
Baby bottles n=6	0.012	0.001	-	0.001	0.003	-	-
	1.400	0.500	-	1.100	0.008	0.001	0.063
	1.773	-	-	-	-	-	-
	0.022	-	-	-	-	-	0.037
	0.915	0.001	-	-	-	-	0.063
	0.003	-	-	-	-	-	0.078

179
 180 The concentrations of bisphenols and phthalates detected in our work did not exceed the
 181 limit of 0.6 mg BPA/kg accepted by the European Union (European Community 2004).

182 All samples analyzed had at least two substances classified as endocrines disruptors. PA
183 and DOP were the compounds most frequently found, followed by BPA. Our results
184 demonstrate the risk of human exposure to endocrine disruptors by migration from food
185 containers.

186

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